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2 **Title:** Do long-distance migratory birds track their niche through seasons?

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4 **Authors:** Damaris Zurell<sup>1,\*</sup>, Laure Gallien<sup>2</sup>, Catherine H. Graham<sup>3</sup>, and Niklaus E.  
5 Zimmermann<sup>1,4</sup>

6 **Affiliation:**

7 <sup>1</sup> Dynamic Macroecology, Dept. Landscape Dynamics, Swiss Federal Research Institute  
8 WSL, Zürcherstrasse 111, CH-8903 Birmensdorf, Switzerland

9 <sup>2</sup> Centre for Invasion Biology, Department of Botany & Zoology, Stellenbosch University,  
10 7602 Matieland, South Africa.

11 <sup>3</sup> Dept. Biodiversity and Conservation, Swiss Federal Research Institute WSL,  
12 Zürcherstrasse 111, CH-8903 Birmensdorf, Switzerland

13 <sup>4</sup> Swiss Federal Institute of Technology ETH, Dept. of Environmental Systems Science,  
14 CH-8092 Zürich, Switzerland

15 **\*Corresponding author:** Damaris Zurell, Email: damaris.zurell@wsl.ch; Tel: +41 44 739  
16 23 70; ORCID ID: 0000-0002-4628-3558

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19

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## 26 **Data accessibility**

27 All data are publicly available; bird range maps at [www.birdlife.org](http://www.birdlife.org), climate data at  
28 [www.worldclim.org](http://www.worldclim.org), bird trait data at

29 <http://dx.doi.org/10.6084/m9.figshare.c.3306933>, bird phylogenetic data at  
30 [www.birdtree.org](http://www.birdtree.org), NDVI at <http://ecocast.arc.nasa.gov/data/pub/gimms/3g.v0/>, and  
31 vegetation cover at <http://glcf.umd.edu/data/landcover/data.shtml>. The species-  
32 specific values for niche overlap, niche tracking and traits are contained in Appendix S2.

## 33 **Abstract**

### 34 **Aim**

35 Seasonal migration by animals is an extensively studied, global phenomenon. Yet, we  
36 still lack a general understanding whether migrants track their niche between summer  
37 and winter ranges and the mechanisms influencing this behaviour. Here, we assessed  
38 the degree of seasonal niche tracking in Holarctic long-distance migratory birds ( $n=717$ ;  
39 excluding very rare species) and evaluate the influence of biogeographic (regional and  
40 range characteristics) and ecological (trophic) factors on tracking.

### 41 **Location**

42 Global.

### 43 **Taxon**

44 Birds.

### 45 **Methods**

46 We calculated seasonal niche overlap by means of ordination, and estimated the degree  
47 of niche tracking using similarity tests. Niche tracking was evaluated for two different  
48 environmental predictor sets: climate and vegetation productivity (reflecting resource  
49 selection) versus climate and land cover (reflecting habitat choice). Multivariate  
50 phylogenetic regression was used to evaluate effects of biogeographic and ecological  
51 traits on niche tracking.

### 52 **Results**

53 We found significant niche tracking in 65-95 % of species with a higher proportion of  
54 species significantly tracking climate and land cover compared to climate and vegetation  
55 productivity. Traits explained 12-18% of the variance in niche tracking with strong  
56 regional differences, a negative effect of migration distance and positive effects of range  
57 size on niche tracking. The effects of niche breadth and trophic traits were less  
58 pronounced and varied between environmental predictor sets.

### 59 **Main conclusions**

60 Our results indicate that at coarse spatial resolution, long-distance migratory species  
61 tend to track their niche and select largely similar environments through seasons.  
62 Stronger niche tracking of land cover could reflect conservatism in habitat selection  
63 across seasons, for example for foraging and roosting. This conservatism towards land

64 cover should be considered when making predictions to future environments. A better  
65 understanding of the factors that constrain seasonal range limits will be critical for  
66 predicting how migration patterns could respond to future environmental changes.

67

68 **Keywords:** animal migration, environmental niche, functional traits, geographic range,  
69 niche breadth, niche tracking, species distribution modelling

70

## 71 Introduction

72 Animal migration is a fascinating, global phenomenon with billions of individuals  
73 travelling between disjunct regions of the world on a regular basis (Fryxell & Sinclair,  
74 1988; Alerstam *et al.*, 2003; Milner-Gulland *et al.*, 2011; Bauer & Hoye, 2014). These  
75 movements likely occur because migrants exploit seasonal resource abundance peaks  
76 (Fryxell & Sinclair, 1988; Tellería & Pérez-Tris, 2003; Somveille *et al.*, 2015; Gómez *et al.*,  
77 2016). A yet underexplored issue is to what extent migratory animals seek comparable  
78 environments throughout the year and thus to what extent they occupy similar niches in  
79 their winter and summer ranges (Milner-Gulland *et al.*, 2011). Birds constitute the best-  
80 studied group of migratory animals. Several studies suggest that not all migratory birds  
81 track their environmental niche between seasons equally well (Boucher-Lalonde *et al.*,  
82 2014; Laube *et al.*, 2015), and some birds have been found to instead switch their  
83 environmental niche between seasons (Joseph & Stockwell, 2000; Martínez-Meyer *et al.*,  
84 2004; Nakazawa *et al.*, 2004). Yet, little is known about the relative proportion of *niche*  
85 *tracking vs. niche switching* strategies in migratory animals and the species  
86 characteristics or seasonal range features that determine these behaviours (Martínez-  
87 Meyer *et al.*, 2004; Nakazawa *et al.*, 2004; Laube *et al.*, 2015; Gómez *et al.*, 2016). Here,  
88 we quantify the degree to which long-distance migratory birds of the Holarctic track  
89 their niches between seasons, and we investigate the role of geographic and ecological  
90 factors in determining the degree to which species track their niches. This knowledge is  
91 critical for predicting how and why species migratory strategies could be influenced by  
92 climate change, and for guiding future monitoring and modelling efforts aimed at the  
93 conservation of migratory animals.

94 An animal's capacity to track suitable environments between seasons (i.e. its  
95 environmental niche) may be affected by biogeographic factors, such as geographic  
96 origin and range size, and by ecological factors, such as body mass and diet specificity,  
97 resulting in a number of testable hypotheses. For instance, the region of origin may  
98 affect a species' capability to track its niche because broad geographic zones differ in  
99 available landmass and spatial arrangement of suitable habitats and climates. This may  
100 all translate into different migration distances, or barriers that stop or funnel migration.  
101 In addition, the geographic position and size of the breeding and overwintering ranges  
102 may affect the distance travelled between those geographic areas (Laube *et al.*, 2015;  
103 Gómez *et al.*, 2016). As long-distance migration is energetically costly (Kranstauber *et al.*  
104 *et al.*, 2015), migrants have to balance the benefits of niche tracking against travel  
105 distances and movement costs (Laube *et al.*, 2015; Gómez *et al.*, 2016). It may thus be  
106 expected that a species' ability to track suitable environments decreases with increasing  
107 breeding to overwintering range distances. Correspondingly, we hypothesize reduced

108 niche tracking (or even niche switching) for species breeding in high-latitudes (Gómez *et al.*, 2016), simply because the cold biomes of high latitudes in the Northern hemisphere  
109 are further apart from their analogous counterparts in the Southern hemisphere than  
110 warmer regions closer to the equator. Furthermore, seasonal niche tracking may  
111 increase with environmental niche breadth or range size of the species (Gómez *et al.*,  
112 2016). Also, we expect the degree of niche tracking to be influenced by ecological and  
113 behavioural characteristics. For example, as flight is energetically more costly in large  
114 birds (Pennycuick, 1969; Hedenström, 1993), we may expect a negative correlation  
115 between the degree of niche tracking and species body mass (Watanabe, 2016), at least  
116 for migrants employing an active flight mode. Diet specificity may also affect species  
117 niche tracking behaviour, as generalists are by definition more flexible than specialist  
118 species regarding required resources.  
119

120 Our objectives were (i) to estimate seasonal niche tracking in long-distance migratory  
121 birds of the Holarctic, and (ii) to test which species characteristics (biogeographic and  
122 ecological attributes) explain variation in the degree of niche tracking between species.  
123 Although migration also occurs in the tropics, we here concentrate on long-distance  
124 migrants breeding in the Holarctic ( $n=717$ , excluding the very rare species), because  
125 these show distinct latitudinal migration patterns and follow comparable resource  
126 pulses triggered by northern-hemisphere summer and winter seasons (Newton, 2007).  
127 First, we estimated birds' breeding (May-July) and overwintering (November-January)  
128 niches using ordinations of environmental data (Broennimann *et al.*, 2012). Second, we  
129 calculated niche overlap between seasons and, third, we estimated niche tracking by  
130 similarity tests that evaluate whether niche overlap was higher than expected by chance  
131 given the environmental conditions. Niche tracking was evaluated for two different  
132 predictor sets: climate and vegetation productivity and climate and land cover. We  
133 assumed that climate is an important determinant of the broad-scale niche of migratory  
134 birds and thus included it in both predictor sets. However, the environmental niche of a  
135 species results from a complex combination of multiple components, including also local  
136 food resource availability as well as habitat choice for foraging and shelter in addition to  
137 climate (Guisan & Thuiller, 2005; Soberón, 2007). Hence, we evaluated whether  
138 migrants would track resource availability (approximated by vegetation productivity)  
139 and habitat (approximated by proportional land cover) across seasons in combination  
140 with climate. Finally, we evaluated our hypothesized relationships between species'  
141 degree of niche tracking and biogeographic and ecological factors (Table 1). To account  
142 for non-independence between species and their traits, species phylogenetic relatedness  
143 was controlled for in all statistical analyses.

## 144 **Tables**

145  
146  
147

**Table 1. Potential biogeographic and ecological factors affecting niche-tracking ability of long-distance migratory birds that were tested in this study. Potential effects of each predictor on species niche-tracking abilities can be positive (+), negative (-), or mixed (+/-).**

Category	Predictors	Hypothesised effect	Hypotheses
Region	<ul style="list-style-type: none"> <li>Region of origin (i.e. Nearctic, Western and Eastern Palearctic; as categorical variables or as breeding longitude)</li> </ul>	+/-	The shape of continents leads to (i) different spatial arrangement of suitable environments resulting in shorter or longer migration distances and, thus, differences in migration costs required for complete niche tracking, or (ii) to differences in the occurrence of migration barriers.
Range position	<ul style="list-style-type: none"> <li>Distance between breeding and overwintering ranges</li> <li>Breeding latitude</li> </ul>	-	<p>Niche tracking ability decreases with migration distance because of increasing energetic and temporal constraints of migration.</p> <p>Species breeding in colder regions (i.e. high-latitude) need to fly longer distances than species breeding in warmer regions (i.e. low latitudes) to reach comparable biomes in summer and winter, which results in a lower degree of niche tracking.</p>
Range size	<ul style="list-style-type: none"> <li>Range size</li> <li>Niche breadth</li> </ul>	+	<p>Species with large range sizes have a higher probability of finding comparative environments in opposing seasons.</p> <p>Broad environmental tolerances lead to a higher probability of finding comparative environments in opposing seasons.</p>



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Ecological traits	• Body mass	-	Body mass increases energetic cost of migration (particularly in birds employing active flight strategies), resulting in lower degree of niche tracking.
	• Diet	+/-	Diet specificity will increase niche tracking if the resource correlates well with climate and vegetation (e.g. fruits), or will decrease niche tracking if the resource does not correlate well with environment (e.g. vertebrates).

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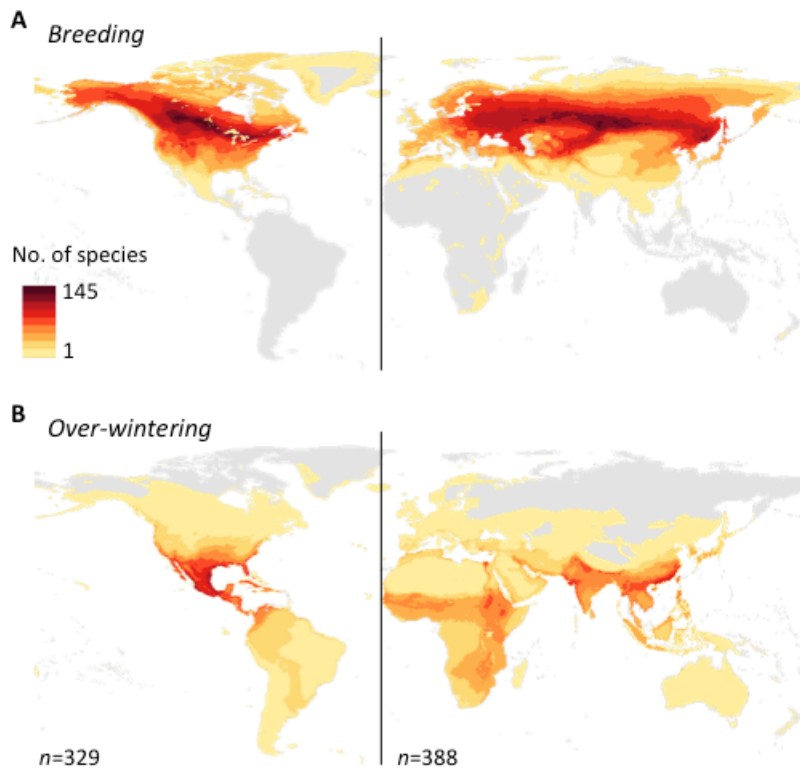
149

150 **Methods**

151 **Species and environmental data**

152 Information on summer breeding and wintering ranges were derived from a global  
 153 dataset of the world's bird species distributions (BirdLife International & NatureServe,  
 154 2014). We checked all synonyms, matched species names with phylogenetic and trait  
 155 data sets (Jetz *et al.*, 2012; Wilman *et al.*, 2014), and excluded species whose taxonomic  
 156 classification changed recently in one of the above databases. We applied a number of  
 157 filtering steps to identify long-distance migrants of the Holarctic: (1) species had distinct  
 158 summer breeding and wintering ranges; (2) the summer breeding range centroid was  
 159 North of 30°N; and (3) the minimum latitudinal distance between summer and winter  
 160 range centroids was 10° (ca. 1100 km). (4) We gridded the summer and winter range  
 161 polygons at a 0.5° resolution, and considered only those species that had at least 40  
 162 presences in their summer and 40 presences in their winter ranges. A species was  
 163 considered as present if its range polygon covered the centre of the gridded cell. Using  
 164 these filtering steps, we selected  $n=717$  extant long-distance Holarctic migratory birds  
 165 for subsequent analyses (Fig. 1). Environmental niches of Nearctic and Palearctic  
 166 breeding birds were analysed separately to control for biogeographic history (329  
 167 Nearctic migrants breeding west of 18°W, and 388 Palearctic migrants).

168



169

170 **Figure 1. Species richness of Holarctic long-distance avian migrants in summer breeding (A) and wintering**  
 171 **(B) ranges. The solid line marks the divide (LON 18° W) between Nearctic (n=329) and Palearctic breeders**  
 172 **(n=388).**

173

174 We considered two different environmental predictor sets to analyse niche tracking, (i)  
 175 climate + NDVI and (ii) climate + land cover. Climate at 0.5° resolution was represented  
 176 by mean temperature and total precipitation during the summer breeding (May-July)  
 177 and wintering season (November-January) and was derived from WorldClim over the  
 178 period 1960-1990 (Hijmans *et al.*, 2005). Similarly, we calculated the mean normalized  
 179 differenced vegetation index (NDVI) separately for each season (1982-2000; GIMMS  
 180 AVHRR Global NDVI; Pinzon & Tucker, 2014). Proportional land cover at 0.5° resolution  
 181 was calculated from the 1km UMD Land Cover Classification (Hansen *et al.*, 2000) and  
 182 was constant throughout the year. We aggregated the original land cover classes into  
 183 seven broad categories (proportional cover of water, forest, shrubland, grassland,  
 184 cropland, bare ground, urban per 0.5° grid cell; SI Appendix, Table S1) for our analyses.  
 185 We verified the appropriateness of the predictor sets for describing summer and winter  
 186 ranges using statistical species distribution models (Guisan & Zimmermann, 2000),  
 187 which showed good to excellent predictive performance (area under the receiver  
 188 operating characteristic curve AUC > 0.7 (Hosmer & Lemeshow, 2000) for all species) in  
 189 repeated (n=3) 70:30 split-sample tests for both seasons and both predictor sets (mean

190 AUC  $\pm$  SD for climate + NDVI: summer  $0.96\pm 0.03$ , winter  $0.98\pm 0.02$ ; climate + land cover:  
191 summer  $0.92\pm 0.02$ , winter  $0.92\pm 0.02$ ).

## 192 Niche tracking

193 Our niche tracking analyses were based on the protocols described in Broennimann *et al.*  
194 *(2012)* and provided in R package *ecospat* (Broennimann *et al.*, 2016), which involves  
195 three basic steps: (1) calculating density of presences and available environment along  
196 niche axes for summer breeding and winter ranges, (2) calculating niche overlap  
197 between summer breeding and winter ranges along the niche axes while correcting for  
198 overall environmental availability, and (3) calculating niche tracking by testing for  
199 significant deviation of niche overlap from random expectation using similarity tests.

200 For (1), we calculated the density of species' presences and of available environmental  
201 factors along the axes of a principal components analysis that summarises the main  
202 environmental variation. The original algorithms for niche overlap analyses  
203 (Broennimann *et al.*, 2012; Broennimann *et al.*, 2016) included only one- and two-  
204 dimensional kernel density estimators, and thus only allowed consideration of two niche  
205 axes. Here, we extended those algorithms by a multivariate kernel density smoother to  
206 allow consideration of up to four niche axes (codes provided in SI Appendix). We  
207 estimated species' niches for two different predictor sets: (i) climate + NDVI, and (ii)  
208 climate + land cover. For the first predictor set, all environmental niche axes (three  
209 variables) could be included in the kernel estimation, while for the second predictor set  
210 (nine variables) only the first four PCA axes could be included. These axes explained  
211 65% and 68% of the environmental variation in Palearctic and Nearctic, respectively (SI  
212 Appendix, Figs. S1).

213 For (2), we calculated niche overlap between summer breeding and winter ranges along  
214 the chosen niche axes using Schoener's  $D$  metric (Schoener, 1968) while correcting for  
215 differences in relative availability of environments by dividing the density of presences  
216 by the density of the environment from the entire environmental space from summer  
217 and winter ranges (Broennimann *et al.*, 2012).  $D$  varies between 0 (no overlap) and 1  
218 (complete overlap).

219 For (3), we used similarity tests to quantify niche tracking. This was done by comparing  
220 the observed niche overlap metric  $D$  between summer and winter niches against a  
221 simulated niche overlap  $D$  metric. To calculate the simulated  $D$ , the entire observed  
222 density of presences of one range was shifted randomly in PCA space. This was done by  
223 randomly selecting the simulated niche centre (the centre of density of presences)  
224 within available environments in the PCA space (Broennimann *et al.*, 2012). We

225 repeated these permutations 200 times (100 permutations where the niche of the  
226 summer range was assumed to be the observed niche while the niche of the winter  
227 range was shifted, and another 100 permutations, where the niche of the winter range  
228 was assumed to be the observed niche while the niche of the summer range was  
229 shifted). For each permutation, we recorded simulated  $D$ , which was defined as the  
230 niche overlap  $D$  between simulated and observed niche. Then, over all 200 permutations  
231 we calculated the standardized effect size  $SES_D = (\text{observed } D - \text{mean of simulated } D) / \text{standard deviation of simulated } D$ . Significant niche switching was identified for  $SES_D$   
232 smaller than -1.64 (i.e. the 5% percentile of a Normal distribution, indicating that  
233 seasonal niches are more dissimilar than expected by chance), while significant niche  
234 tracking was identified for  $SES_D$  larger than 1.64 (i.e. the 95% percentile of a Normal  
235 distribution, indicating that seasonal niches are more similar than expected by chance).  
236

237 Throughout the text, we refer to the raw measure  $D$  as niche overlap, and to  $SES_D$  as  
238 niche tracking.

### 239 Trait analyses

240 Trait information were extracted from Wilman *et al.* (2014) and phylogenetic  
241 information from Jetz *et al.* (2012). We included the following functional traits: body  
242 mass and dominant diet type (invertebrates; vertebrates incl. fish and carrion; fruits and  
243 nectar; plants and seeds); all species that could not be assigned to one dominant diet  
244 category were classified as omnivores (Wilman *et al.*, 2014). Additionally, we considered  
245 breeding longitude and breeding latitude (calculated from summer range centroids), the  
246 latitudinal distance between summer and winter range centroids, summer and winter  
247 range size, and niche breadth in trait analyses. Niche breadth was quantified following  
248 Laube *et al.* (2015), by calculating the Shannon index from environmental occupancy  
249 (the density of species occurrences in PCA space divided by the density of available  
250 environment; cf. niche tracking analyses above), which takes into account the number of  
251 occupied grid cells and the evenness in occupancy. From the literature, we compiled  
252 information on flight mode, distinguishing between the broad categories of mainly  
253 active (flapping) and mainly passive (soaring) flight (SI Appendix, Table S1).

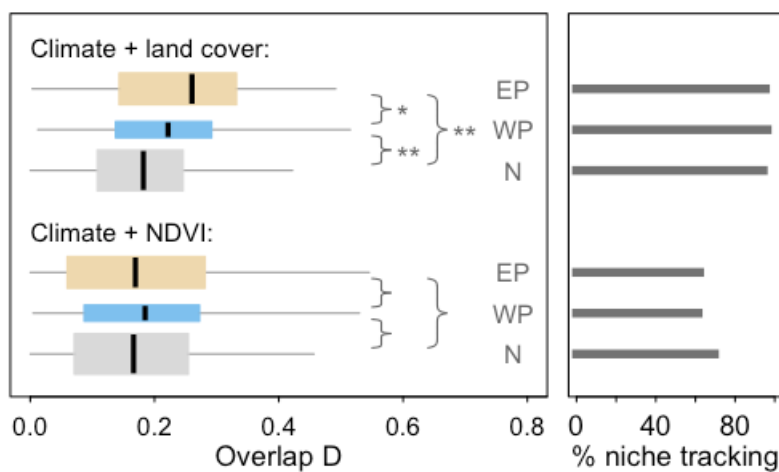
254 We used phylogenetic regression to estimate the effect of the biogeographic and  
255 ecological traits on niche tracking (Table 1) while controlling for non-independence  
256 between species and their traits due to phylogenetic relatedness (Paradis & Claude,  
257 2002) using the R package *phylolm* (Ho & Ane, 2014). For longitude, we included linear  
258 and quadratic terms to test for non-linear effects. Parsimonious models were identified  
259 using AIC-based stepwise variable selection. Variable importance was calculated based  
260 on an algorithm where each variable was permuted randomly ( $n=99$ ) in order to

261 mimic the absence of the variable in the model (Strobl *et al.*, 2007). Then, importance  
 262 was estimated according to the difference in deviance explained by the models with  
 263 and without permutation. In the main text, we present trait analyses including all long-  
 264 distance migrants (n=717). Additionally, we repeated trait analyses including only  
 265 those migrants employing active flight strategies (n=637), to test for potentially  
 266 confounding effects of different energetic costs of active vs. passive flight on niche  
 267 tracking. Results for active flying migrants are presented in the SI appendix.

268 All analyses were conducted using R version 3.2.3 (R Core Team, 2015).

## 269 Results

270 Estimated seasonal niche overlap  $D$  for climate + NDVI ranged 0.00-0.56 (mean=0.18,  
 271 sd=0.12), and for climate + land cover ranged 0.00-0.56 (mean=0.21, sd=0.11). Niche  
 272 overlap  $D$  estimated for climate + land cover showed significant differences across  
 273 regions, with Eastern and Western Palearctic migrants (east and west of 65°E) showing  
 274 larger  $D$  than Nearctic migrants (west of 18°W). No such differences were apparent for  
 275 climate + NDVI niche overlap comparison (Fig. 2). Overall, niche overlap for the two  
 276 different predictor sets (climate + NDVI vs. climate + land cover) was significantly  
 277 correlated (Pearson correlation coefficient  $r=0.67$ ,  $t=24.22$ ,  $p<0.001$ )(Fig. 3). The  
 278 amount of niche tracking estimated from similarity tests differed between predictor sets  
 279 with 65 % of species significantly tracking their climate + NDVI niche, and 95 % of  
 280 species significantly tracking their climate + land cover niche (Figs. 2-3). We identified  
 281 no case of significant niche switching. In addition, niche overlap  $D$  was significantly  
 282 larger for migrants employing mainly passive flight and also the proportion of niche  
 283 tracking migrants was slightly larger for migrants employing mainly passive flight (SI  
 284 Appendix, Fig. S2).

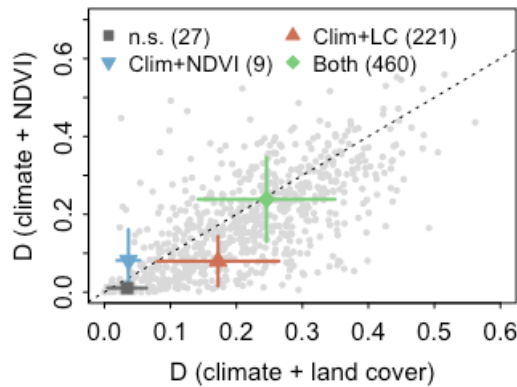


285

286 **Figure 2. Niche overlap  $D$  between summer breeding and winter ranges (left) and proportion of species**  
 287 **significantly tracking their seasonal niche (right). We distinguish long-distance migratory birds breeding in**

288 Nearctic (N), Western Palearctic (WP) and Eastern Palearctic (EP). Niche overlap  $D$  was estimated along niche  
 289 axes obtained from PCA for the two predictor sets climate + NDVI, and climate + land cover; outliers are not  
 290 displayed. Asterisks indicate significant differences following a two-tailed Wilcoxon rank sum test (\*\*  $p <$   
 291  $0.01$ ; \*  $p < 0.05$ ). The width of the boxplots is proportional to the species numbers within geographic regions  
 292 ( $N n=329$ ;  $WP n=132$ ;  $EP n=256$ ). The barplot (right) indicate results from the similarity tests where the niche  
 293 (in PCA space) during summer is compared to random niches during winter and vice versa.

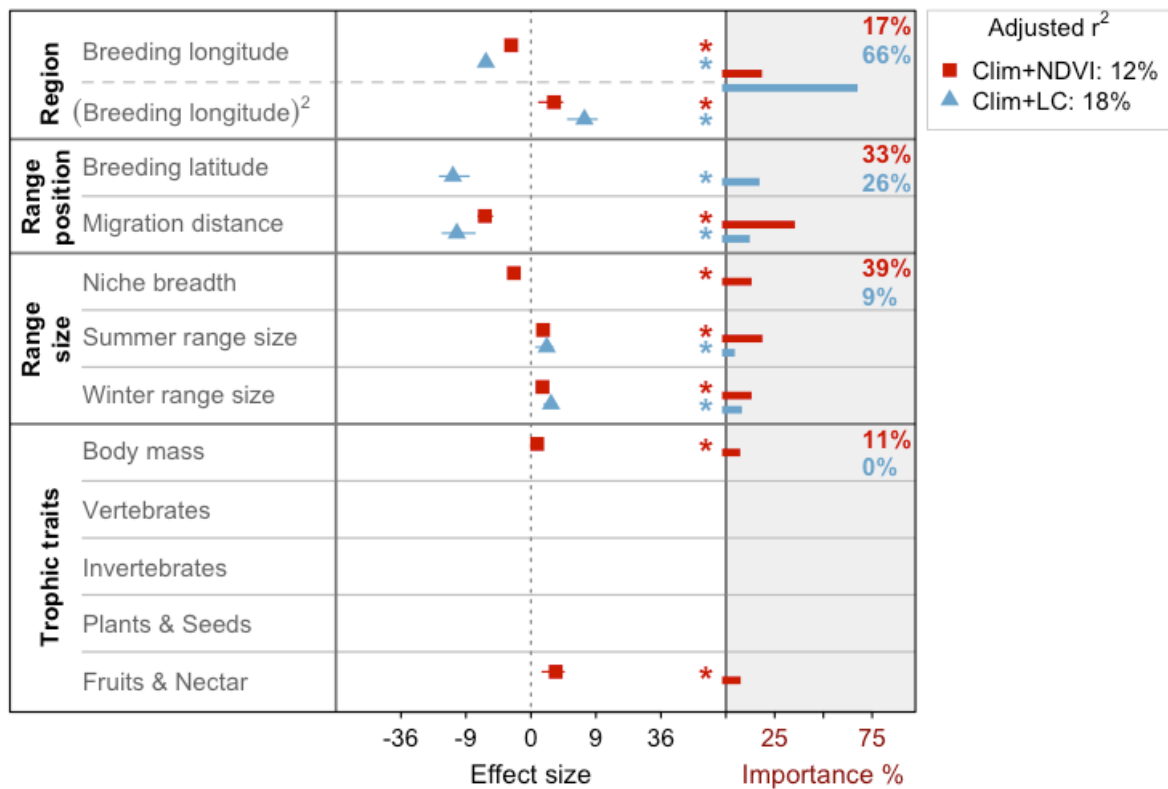
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295

296 **Figure 3.** Correlation between estimated niche overlap  $D$  for the two predictor sets (climate + NDVI, climate +  
 297 land cover). Grey points show overlap  $D$  for all species ( $n=717$ ). Coloured points show the mean  $D$  (with  
 298 standard deviation) for species showing no significant niche tracking (darkgrey), species showing significant  
 299 niche tracking only for climate + NDVI (blue), only for climate + land cover (darkred), and species showing  
 300 significant niche tracking for both predictor sets (green) with the number of species shown in brackets.

301 The considered traits explained 12 % and 18 % of the variance in niche tracking ( $SES_D$ )  
 302 for the predictor sets climate + NDVI and climate + land cover, respectively. The  
 303 importance of biogeographic and ecological traits in predicting niche tracking  
 304 behaviours varied across traits (Fig. 4). The breeding region was most important for  
 305 explaining niche tracking of climate + land cover while factors related to range size and  
 306 range position were most important for explaining niche tracking of climate + NDVI.  
 307 Among the traits related to range position, migration distance had a consistent negative  
 308 effect on niche tracking, while breeding latitude had a significant negative effect only for  
 309 climate + land cover niche tracking. Summer and winter range sizes showed significant  
 310 positive effects on niche tracking. Niche breadth showed a significant negative effect for  
 311 climate + NDVI niche tracking. Trophic traits were the least important predictors of  
 312 climate + NDVI niche tracking with significant positive effects of body mass and  
 313 frugivorous diet. By contrast, trophic traits had no effect on climate + land cover niche  
 314 tracking. When considering active flight only, the effects of the considered biogeographic  
 315 and ecological traits on niche tracking were largely consistent except for the diet traits  
 316 that had no effect in any of the environmental predictor sets (SI Appendix, Fig. S3).



317

318 **Figure 4. Relationship of species niche tracking with biogeographic and ecological factors.** Niche tracking is  
 319 given by the standardised effect size  $SES_D$  estimated from niche similarity tests. To explain  $SES_D$  values from  
 320 traits, multivariate models were estimated by phylogenetic regression using AIC-based stepwise variable  
 321 selection ( $n=717$ ). Asterisks indicate significant coefficients ( $p < 0.05$ ). Bars indicate importance of each  
 322 variable selected in the final model (note that linear and quadratic terms of longitude are summarised into a  
 323 single importance value), and given percentages sum the variable importance for the four different categories  
 324 of biogeographic and ecological factors (cf. Table 1). Overall explained variance is given by the adjusted  $r^2$  of  
 325 the multivariate model.

326

## 327 Discussion

328 Every year, billions of animals migrate large distances between their breeding and  
 329 overwintering ranges. An outstanding question is whether these migratory species  
 330 select largely similar environments through seasons. Using a cross-continental analysis  
 331 on all but the very rare long-distance migratory birds breeding in the Holarctic, we  
 332 demonstrate that most long-distance migrants do significantly track their broad-scale  
 333 environmental niches through seasons indicating strong niche conservatism at large  
 334 spatial scales. Our analyses rely on coarse-scale range maps of species and should be  
 335 interpreted cautiously because of potential range size overestimation especially in  
 336 winter ranges. Nevertheless, our results indicate that factors related to region, range  
 337 position and range size seem more important in determining seasonal niche tracking at



338 large spatial scales than ecological factors related to trophic traits. A better  
339 understanding of the factors that constrain seasonal range limits will be crucial for  
340 improved prediction of how migration patterns could respond to future climate changes  
341 as such changes have the potential to affect ecosystem functions worldwide (Bauer &  
342 Hoyer, 2014).

343 We explored seasonal niche tracking at broad spatial scales and using two different  
344 predictor sets. Overall, our results indicate that long-distance migratory birds largely  
345 track their niche through seasons with more than 65 % of the species selecting more  
346 similar environments than expected by chance. This is well in line with previous findings  
347 on new world warblers that identified significant tracking of climatic niches in 73 % of  
348 migrants (Gómez *et al.*, 2016). Other studies had reported niche trackers and niche  
349 switchers but are not directly comparable as their niche overlap measures were based  
350 on geographic projections (e.g., Martínez-Meyer *et al.*, 2004; Nakazawa *et al.*, 2004)  
351 which could be biased by differing spatial structure of environments in the opposing  
352 seasons (Broennimann *et al.*, 2012), or because niche tracking and switching was not  
353 tested explicitly by similarity tests (e.g., Laube *et al.*, 2015). Here, we estimated stronger  
354 niche tracking for climate and land cover compared to climate and vegetation  
355 productivity. This indicates that at coarse spatial resolution species tend to select  
356 similar environments, but are slightly less conservative in respect to vegetation  
357 productivity than to land cover. This seems to imply strong conservatism in important  
358 attributes such as in the foraging niche and in other behaviours such as roosting that are  
359 better reflected in land cover types than in vegetation productivity. Stronger seasonal  
360 niche conservatism for land cover also suggests that land use and land cover are  
361 important predictors for projecting species response to global change, while studies  
362 focussing on breeding ranges alone mostly found climate to be a more important  
363 predictor (Hockey *et al.*, 2011; Howard *et al.*, 2015).

364 We explored four main hypotheses for how biogeographic and ecological characteristics  
365 influence the degree of seasonal niche tracking (Table 1). The considered traits  
366 explained 12-18 % of the variation in seasonal niche tracking, indicating that large-scale  
367 niche tracking is at least to some extent related to certain species characteristics.  
368 Overall, important generalities emerged. First, we found strong support for our  
369 hypothesis that geographic origin affects seasonal niche tracking, which was true for  
370 both predictor sets although much more pronounced for climate and land cover niche  
371 tracking. As we used the standardised effect sizes from similarity tests to assess niche  
372 tracking rather than the raw niche overlap values, the results are not biased by  
373 availability of environment but must be caused by other geographic features. For  
374 example, differences in migration routes could affect niche tracking through migration



375 costs in terms of energy and time (Hedenström, 1993; Wikelski *et al.*, 2003) and  
376 atmospheric conditions among others (Sapir *et al.*, 2011; Kranstauber *et al.*, 2015).  
377 Additionally, barriers such as oceans, mountains and deserts may limit migration and  
378 niche tracking abilities differently between continents, for instance by retaining species  
379 in areas of lower habitat suitability and thus reducing the niche tracking abilities.

380 The trait analyses also corroborated our second hypothesis that increased migration  
381 cost due to longer migration distance could reduce niche tracking. Overall, species with  
382 large migration distance are less likely to track their climate and vegetation productivity  
383 or climate and land cover niche through seasons. Previous studies found either no  
384 relationship between migration distance and niche overlap, as was the case for topo-  
385 climatic niches of New World warblers (Gómez *et al.*, 2016), or found a negative  
386 relationship only when considering land cover niche overlap, as was shown for *Sylvia*  
387 warblers (Laube *et al.*, 2015). Our study thus provides new evidence that migratory  
388 distance between breeding and over-wintering grounds may strongly affect niche  
389 tracking. Breeding latitude, on the other hand, showed mixed results. In line with our  
390 initial hypothesis, it had a significant negative effect on climate and land cover niche  
391 tracking. By contrast, breeding latitude had no effect on climate and NDVI niche  
392 tracking. Thus, more northerly breeding grounds could incur higher migration costs and  
393 decrease niche tracking ability in long-distance migrants but, given the inconsistency  
394 between predictor sets, this conclusion cannot be generalized.

395 Third, we found support for increased niche tracking in species with larger range sizes  
396 but no support for increased niche tracking with wider niche breadth. Large range sizes  
397 increased the probability of finding comparable environments in opposing seasons for  
398 both predictor sets. Such a positive effect of range size, after controlling for statistical  
399 bias through similarity tests, could indicate other biotic controls on seasonal niche  
400 tracking. Among the factors that are hypothesized to correlate with range size are traits  
401 such as habitat niche, diet niche, life history traits such as fecundity, dispersal ability,  
402 body size, historical legacies (related both to geographic and evolutionary factors) and  
403 interspecific interactions (Brown *et al.*, 1996; Laube *et al.*, 2013). In our study, we  
404 separately assessed the effects of diet, body size, niche breadth, and controlled for  
405 phylogenetic history (i.e. using phylogenetic regressions). By contrast, dispersal ability,  
406 life history characteristics (fecundity) and interspecific interactions could not be  
407 evaluated, but are interesting candidates for future (population-level) analyses of  
408 seasonal niche tracking (Gómez *et al.*, 2016). For niche tracking of climate and  
409 vegetation productivity, we found an unexpected negative effect of niche breadth. This  
410 effect could be related to the calculation of niche breadth, which considers the total  
411 annual niche of the species. However, the range of temperature and precipitation values

412 experienced in winter seems to be larger than that experienced in summer while no  
413 such effect was found for NDVI (Fig. S4). Thus, the negative effect of niche breadth on  
414 niche tracking of climate and vegetation productivity could actually mirror a low  
415 seasonal overlap in the temperature niche and precipitation niche.

416 Last, we only found weak support for our fourth hypothesis related to trophic traits. In  
417 contrast to our expectation, we found a positive effect of body mass on niche tracking of  
418 climate and vegetation productivity although this effect vanished when considering  
419 active flight only. Thus, overall, our analyses suggest that body mass has no significant  
420 effect on seasonal niche tracking in long-distance migratory birds. Functional traits  
421 related to diet also did not seem to play an important role in niche tracking. We only  
422 found a significant positive effect of frugivorous diet on niche tracking of climate and  
423 vegetation productivity. This result supports our hypothesis that diet specificity will  
424 increase niche tracking if the resource correlates well with climate and vegetation as in  
425 fruits, but seems not to be generalizable across environmental predictor sets. Overall,  
426 seasonal niche tracking at large spatial scales seems to have little functional signals but  
427 is primarily determined by geographic and potentially range size related effects.

428 These results are contingent on broad-scale environmental niches inferred from range  
429 maps. Generally, range maps may not be equally precise in all species, may include a  
430 high number of false presences (Hurlbert & Jetz, 2007), and may have better evidential  
431 support in the summer than in the winter ranges (Brown *et al.*, 1996), which could  
432 potentially bias our analyses. Here, we quantified niche overlap in environmental space  
433 using ordination and kernel density estimation, which should reduce the problem of  
434 spatial bias in occurrence records (Broennimann *et al.*, 2012; Gómez *et al.*, 2016). Also,  
435 we tried to reduce potential bias by controlling for range size effects using similarity  
436 tests. For future analyses, it would be desirable to quantify seasonal niche tracking at  
437 finer spatial scales and possibly at population-level using occurrence records that  
438 become increasingly available, for example through eBird data portal ([www.ebird.org](http://www.ebird.org)).  
439 Yet, such citizen science data currently suffer from uneven sampling efforts and size and  
440 from potential spatial biases especially in winter ranges.

441 With increasing loss of biodiversity due to global environmental change, there is a  
442 strong need to forecast the species' responses to these changes (Urban *et al.*, 2016;  
443 Zurell *et al.*, 2016). This is especially challenging in long-distance avian migrants, as they  
444 are affected by environmental conditions in different parts of the world. Our results  
445 identify two important questions for migratory bird responses to global change. (i)  
446 Which factors limit the range of a species, and how plastic are species seasonal niche  
447 requirements? The large proportion of significant niche tracking for climate and  
448 vegetation productivity and in particular for climate and land cover indicate that abiotic

449 environmental variables other than climate also may be important for predicting species  
450 response to climate change. (ii) How will migration strategies respond to global change?  
451 Our results suggest that migration cost plays an important role in determining the  
452 degree of niche tracking. Thus, if climate and land use change lead to an increased  
453 distance between suitable breeding and overwintering ranges, this could then alter  
454 migration strategies. On the one hand, migration may cease altogether with important  
455 repercussions on biodiversity patterns and ecosystem functioning (Bauer & Hoyer,  
456 2014). On the other hand, migratory species may shift from seasonal niche tracking to  
457 seasonal niche switching with unclear effects on the stability and size of their  
458 populations, making their response to and survival under global change even less  
459 predictable. Long-term observations in migratory population dynamics by means of  
460 telemetry and censuses may help to gain a more comprehensive understanding of the  
461 complex nature of animal migration (Rotics *et al.*, 2016), seasonal niche tracking, and  
462 population development through time. Improving the intensity in surveying distribution  
463 and population sizes is especially important in overwintering ranges for improving our  
464 capability to forecast and conserve migratory animals in the future.

465

## 466 **Supplementary Material**

467 Appendix S1. Supplementary results.

468 Appendix S2. Species data table.

469

470

## 471 **Biosketch**

472 Damaris Zurell is broadly interested in macroecology and biodiversity modelling. Main  
473 topics include environmental change effects on spatio-temporal structuring of species'  
474 ranges, populations and communities, and the evolution of life history characteristics,  
475 especially in migratory species.

476

477 **Author contributions**

478 DZ and NEZ conceived the idea. DZ compiled all data and ran the analyses with the aid of  
479 LG, prepared the figures and wrote the initial draft of the paper. All authors contributed  
480 to interpreting the results and to writing the manuscript.

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